

1 **The Greenest Path: Comparing the Effects of**
2 **Internal and External Costs of Motor Vehicle**
3 **Pollution on Route Choice and Accessibility**

4 Mengying Cui
5 University of Minnesota
6 Department of Civil, Environmental, and Geo- Engineering
7 500 Pillsbury Drive SE
8 Minneapolis, MN 55455 USA
9 cuixx242@umn.edu

10 David Levinson
11 RP Braun-CTS Chair of Transportation Engineering
12 Director of Network, Economics, and Urban Systems Research Group
13 University of Minnesota
14 Department of Civil, Environmental, and Geo- Engineering
15 500 Pillsbury Drive SE
16 Minneapolis, MN 55455 USA
17 dlevinson@umn.edu

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1 **ABSTRACT**

2 On-road emissions are a dominant source of urban air pollution, which damages human health.
3 The "greenest path" is proposed as an alternative pattern of traffic route assignment to minimize
4 the costs of emissions or exposure, pursues an environmentally optimal. The framework of a link-
5 based emission cost analysis is built for both internal and external environmental costs and applied
6 to the road network of the Twin Cities Metropolitan area based on the EPA MOVES model. The
7 greenest (internal/external) path is skimmed for all OD pairs to compare the work trip flows on
8 the roads and accessibility distribution. It is shown that the emission cost that travelers impose on
9 others is greater than which they bear. Considering only external emissions costs thus produces
10 a lower accessibility than considering only internal emissions costs. This research contributes to
11 understanding the full cost of travel.

1 INTRODUCTION

2 Outdoor urban air pollution is a major risk to health. According to the World Health Organization,
3 urban air pollution is one of the top 15 causes of death globally, and one of the top 10 causes in
4 medium- and high-income countries. Health effects of urban air pollution include respiratory and
5 cardiovascular disease, and adverse birth outcomes. Exposure to high concentration of airborne
6 particle matter (PM) correlates with many adverse respiratory and cardiovascular health problems,
7 as revealed through epidemiological and toxicological studies (e.g. Dockery (1), Pope III et al.
8 (2)).

9 Motor vehicles are a dominant source of urban air pollution. As a result of incomplete
10 combustion of fossil fuels, a number of contaminants are released into the environment, including
11 carbon monoxide, hydrocarbons, smog-forming constituents, and particulate matter (PM).^{1 2}

12 Once emitted into the atmosphere, air pollutants undergo mixing, diffusion, or chemical re-
13 actions, the degree of which depends on background concentration, meteorological and geographi-
14 cal conditions, and other local characteristics. Since exposure to significant levels of contaminants
15 harms human health, regulations on air quality and vehicle emissions are employed in many coun-
16 tries. In the United States metropolitan areas must certify that transportation plans conform to air
17 quality standards which set maximum allowable levels of criteria pollutants (failure to do so results
18 in suspension of federal highway funds).

19 On-road emissions affect vegetation, materials, aquatic ecosystems, visibility, and climate
20 as well as human health (11). Notably, damage to human health due to air pollution is the most eco-
21 nomically expensive element. Small and Kazimi (12) found that particulate matter is the primary
22 cause of mortality and morbidity.

23 The social cost of pollution depends on highly localized intake and health effects. There
24 remains uncertainty about the operating characteristics of vehicles which minimize pollution, as
25 this depends very much on the nature of the vehicle and the driving conditions. There is less
26 uncertainty about exposure, where pollution intake occurs, as traffic counts and individual travel
27 routes are readily employed.

28 If travelers took alternative routes with reduced pollution generation or intake (exposure),
29 the economic measure of environmental externalities of travel would be lower (13, 14). The full
30 private or internal costs depend on the relative value of reduced pollution intake vs. increased travel
31 time and distance. The full external costs depends on the value of reduced pollution emissions and

¹The switch to hybrid, and ultimately electric vehicles, improves the situation, particularly tailpipe emissions, but does not eliminate the pollution problem. With full electrification of the fleet decades away, the need to mitigate the effects of automobile pollution remains especially salient. Further, all vehicles also generate particulates where the rubber meets the road, from abrasion processes like tires and brake wear, and road dust resuspension (3, 4, 5). Thus, even with electrification, automobile pollution will not disappear.

²Recent health-effects research highlights the potential importance of very small particles of which vehicles are a major contributor to emissions and exposures. Ultrafine particles (UFP), PM with diameter less than $0.1\mu\text{m}$, have drawn increasing research interest because they are:

1. more toxic to lab animals (6, 7),
2. harder to remove by the lung's macrophage clearance mechanism (8, 9), and
3. able to enter the circulatory system and travel to the heart, brain, and other organs (10).

To date, no criteria pollutant standards have been established for UFPs, and estimates are not produced by standard approaches like the EPA MOVES model. Future research should aim to incorporate UFPs in the kind of analysis presented herein.

1 increased total travel costs. The full costs combine the internal and external costs.

2 Some pollution effects would be larger due to increased travel distance or different perfor-
3 mance characteristics, so designing such routes requires great care. We find traffic routing patterns
4 which minimize the costs of air pollution emissions and exposure. These Environmentally User
5 Optimal (EUO), considering exposure, and Environmentally System Optimal (ESO), considering
6 emissions, traffic route assignments will aim to find routes, hereafter called *greenest paths*, that
7 minimize intake or emissions respectively for internal or external cost versions, subject to travelers
8 engaging in the same sets of activities at the same location.³

9 This study first measures the internal and external emission cost for each link segment
10 on the road network based on the EPA MOVES model and provide a link-based emission cost
11 analysis. This paper estimates the average and incremental pollution produced by travelers along
12 various routes. This paper then estimates the traveler pollution intake along various travel paths as
13 a function of endogenous traffic levels.

14 The greenest path was found for all the OD pairs in the seven county Minneapolis - St.
15 Paul (Twin Cities) Metro Area. These paths were then aggregated into an accessibility measure.
16 The resulting accessibility measures are compared with traditional measures (travel time based
17 accessibility). We then analyze the accessibility loss of using the greenest path, which indicates
18 the penalties of it to pursue the minimum emission cost expressed by accessibility. Moreover, we
19 also estimate traffic that would take the greenest path for the actual OD pairs in Twin Cities, if that
20 were the selected path for individual travelers.

21 The data, introduction of MOVES, methodology, accessibility measurement and conclu-
22 sion of this research are shown in sections 3-7 respectively.

23 **DATA**

24 Several sources of data used.

25 • **TomTom Data**

26 Speed and road network data were acquired by the research team from the Metropolitan
27 Council, which has licensed the data from TomTom. The speed data were aggregated
28 and processed based on millions of GPS logging and navigation devices (15). The speed
29 data were categorized by road classification, time period, and speed percentile. Depend-
30 ing on 8 Functional Roadway Classifications (FRC), speed data were separated into 4
31 groups, in which the five categories representing the highest level roads, FRC0 to FRC4,
32 were combined as one. This analysis uses the morning peak period travel times and the
33 50th percentile (median) speed on each link for shortest path estimation. The road net-
34 work, a GIS shapefile that contains the geographic information of each link in the Twin
35 Cities, was joined with the speed data, and were used as an input of MOVES to estimate
36 emissions and to search the greenest path and the shortest travel time path by joining the
37 emission or time cost with the road network.

38 • **2010 Transportation Analysis Zone (TAZ) system**

³We recognize that not all intake is equal, and someone who breathes in 10 units of pollution may be more than 10 times worse off than someone who breathes in a unit of pollution, but we anticipate that the existing research cannot fully establish these nonlinear effects in the context we describe given the available data about existing health status.

1 The Transportation Analysis Zone (TAZ) system in Twin Cities was developed by the
2 Metropolitan Council. This data is a polygon shapefile that shows the 2010 TAZ bound-
3 aries, aggregations, and the employment and household information for each TAZ (16).
4 The features in the seven county Metro area were selected and the centroid of each poly-
5 gon was extracted as the origin and destination for accessibility measurement. Its em-
6 ployment data were used to express the number of jobs in each TAZ.

7 • **LEHD Origin-Destination Employment Statistics dataset (LODES7.0)**

8 The Longitudinal Employment Household Dynamics (LEHD) Origin-Destination Em-
9 ployment Statistics dataset (LODES7.0) was acquired from the United States Census
10 Bureau (17). The data contains tables of Workplace Area Characteristics (WAC), Res-
11 idence Area Characteristics (RAC) and Origin and Destination Census blocks for each
12 worker. LODES data was used to estimate the work trip flows on link segments based on
13 the definition of betweenness with the assumption that all workers in the dataset would
14 select the greenest path or the shortest travel time path.

15 • **Pavement Condition Data**

16 Pavement condition data was collected by the research team from Minnesota Department
17 of Transportation (MNDOT). It includes the pavement quality indicator (PQI), ride qual-
18 ity index (RQI), surface rating (SR), and, most relevant for our analysis, truck percentage
19 (P) for highway links in Minnesota from 2000 to 2015. A corresponding shapefile was
20 also provided by MNDOT to locate the data point on the network. Pavement condition
21 data of the Seven County Metropolitan area of Twin Cities were selected and the truck
22 percentage (P) was used to measure the link type source, which is an important input of
23 MOVES.

24 • **AADT**

25 An estimate of Annual Average Daily Traffic (AADT) from MnDOT, which reports
26 AADT from 1992 to 2014. Data from 2007-2014 from the Twin Cities was joined with
27 road network AADT was used as an input of MOVES as well.

28 **POLLUTION ESTIMATION**

29 The Motor Vehicle Emission Simulator (MOVES), developed by United States Environmental Pro-
30 tection Agency (18), is a modeling system that is widely used to estimate air pollutants, greenhouse
31 gases, and air toxics at the level of nation, county, or project. To estimate the air pollutants for each
32 link segment on the scale of network in the metropolitan area, the project level of simulation was
33 conducted for each county in the Twin Cities. The inputs, such as meteorology and fuel type, vary
34 across counties. The results are subsequently combined for the whole road network. Most of the
35 importers for the simulations were set as the defaults specific to time and location except the tables
36 for links and link source types.

37 • **Links**

38 The link table defines the individual roadway links containing the properties as segment
39 length, traffic flow, average speed and road grade. Segment length was directly measured

1 based on the shapefile of the road network. Traffic flow, representing the units of vehicles
 2 per hour, was extracted from the AADT data and a morning peak hour (7AM-9AM)
 3 volume ratio of 6.09 (The number is a average ratio of morning peak hour flow to AADT
 4 based on the data from MnDOT's IRIS traffic database for the Twin Cities (19)).The
 5 average speed for each link was set as the 50th percentile speed in the morning peak hour
 6 of the speed data. The road grade for all the links were set as 0, which could be improved
 7 with Digital Elevation Model Data.

8 • Link Source Type

9 Link source type describes the fraction of the link traffic flow that is represented by
 10 each vehicle type (source type). There is no available link source type data that could be
 11 directly used for each link segment on the Twin Cities road network. Hence, we estimated
 12 the vehicle type fraction based on truck percentage P . For highway link segments, P
 13 was collected from the pavement condition data, which covers different types of trucks,
 14 single unit trucks, buses, combination trucks. Combining the vehicle source type defined
 15 in MOVES, the setting of the vehicle type fraction for highway link segments shows
 16 as Table 1. The truck percentages of other link segments were estimated based on the
 17 regression of highway truck percentage P on AADT ranges, which is shown in Table 2
 18 (20). And the fraction settings followed the rules in Table 1 based on the estimated P as
 19 well.

TABLE 1 : Vehicle Type Fraction Setting for Highway Link Segment

MnDOT use type	Avg Percentage	MOVES use type	Value
Autos, pickups	1-P	21.Passenger Cars	0.698*(1-P)
		31.Passenger Truck	0.292*(1-P)
		11.Motorcycle	0.010*(1-P)
Buses, Trucks w/ Trailers	0.207*P	41. Intercity Bus	0.207*P/3
		42. Transit Bus	0.207*P/3
		43. School Bus	0.207*P/3
Single Unit Truck (SU)	0.505*P	51. Refuse Truck	0.505*P/3
		52. Single-Unit Short-Haul Truck	0.505*P/3
		53. Single-Unit Long-Haul Truck	0.505*P/3
Combination Truck (TST)	0.288*P	61. Combination Short-Haul Truck	0.288*P/2
		62. Combination Long-Haul Truck	0.288*P/2

¹ The fractions of Buses (0.207), Single Unit Truck (0.505), and Combination Trucks (0.288) were based on average statewide vehicle classification (20).

² The fractions of passenger cars (0.698), passenger trucks (0.292) and motorcycle (0.001) were based on the MnDOT Procedure Manual for Forecasting Traffic on Minnesota's Highway Systems (21).

TABLE 2 : Truck Percentage Estimation Based on AADT Range

Variables		Estimate	Std. Error	Sign
Intercept		10.4168	0.251	***
AADT Range	400-1499	0.928	0.2677	***
	1500-7000	-1.1968	0.261	***
	>7000	-2.4428	0.2618	***
R^2		0.049		

1 METHODOLOGY

2 Internal Emission Cost Estimation

3 The internal emission cost for each traveler is defined as the health damage cost borne due to
 4 pollution intake during trips, which highly depends on the concentration of pollutants, travelers'
 5 breathing rate, exposure time and unit damage cost of pollutants (22).

$$Q_{I,p} = \int_0^T B_r * C_p(t) dt \quad (1)$$

6

$$E_{c,I} = \sum_p Q_{I,p} * u_{I,p} \quad (2)$$

7 Where:

8 $Q_{I,p}$ stands for the quantity of intake-emission of pollutant p ,

9 $E_{c,I}$ stands for the internal emission cost,

10 $C_p(t)$ stands for the concentration of pollutant p , which varies with the time,

11 T stands for the total exposure time,

12 B_r stands for the breathing rate,

13 $u_{I,p}$ stands for the unit intake damage cost of pollutant p .

14

15 A link-based intake emission cost estimation ($E_{c,I,i}$) considers the pollution concentration
 16 ($C_{p,i}$) and travel time (T_i) needed for a specific link segment (i).

17 • Concentration by Link Segment

18 Generally, concentration of pollutants could be expressed as air pollutant part per million
 19 by volume ($ppmv$) or milligrams of pollutant per cubic meter of air (mg/m^3). On the
 20 basis of the quantity of emissions from MOVES, concentration could be measured by air
 21 pollution dispersion models (23).

22 Notably, concentrations of nodes (or intersections) on link segments are much higher
 23 than those of edges (or links) since a node could connect many edges and the vehicles
 24 passing through it from different directions affect its concentration. Hence, to measure
 25 the concentration, each link was divided into three parts, including one part representing
 26 the link exclusive of the junction, and two parts representing buffers of the end junctions,
 27 as shown in Figure 1.

28 A buffer was set for each node referring to its boundary (red circles in Figure 1). It was
 29 assumed that, in a windless scenario, the concentration of a node equals the sum of the
 30 concentrations of all edges passing through it, which could be expressed as,

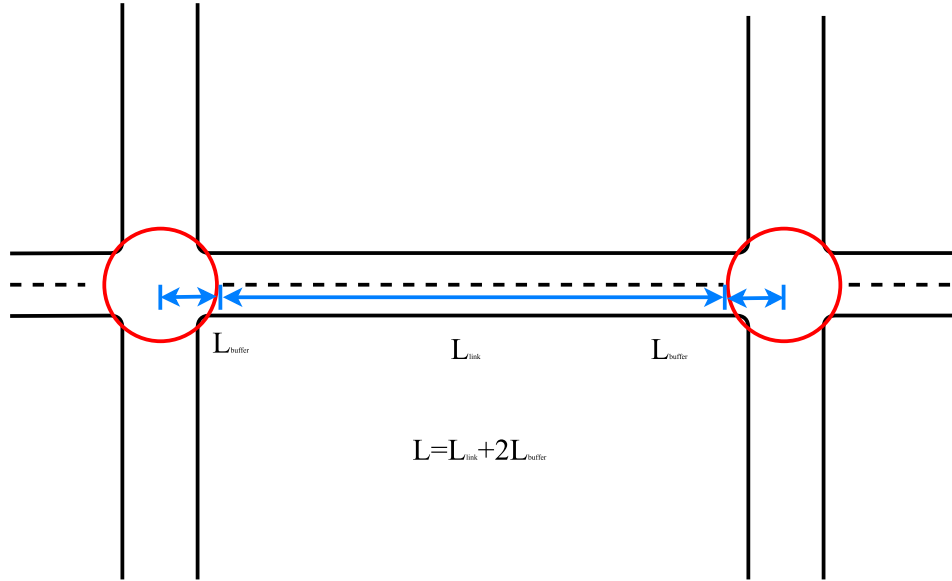


FIGURE 1 : Link-based Concentration Measurement

$$C_{p,n_i} = \sum_{k \in n_i} C_{p,e_k} \quad (3)$$

1 Where:

2 C_{p,n_i} stands for the concentration of pollutant p on a node n_i ,

3 C_{p,e_k} stands for the concentration of pollutant p on link k , which connects with node n_i .

4 The concentration of an edge was assumed to be only affected by the vehicles in its own
5 direction. The concentration of a buffer considers traffic on the link and in the relevant
6 crossing direction.

7 • **Air Pollution Dispersion Model**

8 The windless air pollution dispersion model, in which the speed of wind is 0.4m/s or less,
9 could be expressed as (24),

$$C_{p,i} = \frac{2q_{p,i}}{(2\pi)^{3/2}} \int_{l_1}^{l_2} \frac{1}{\beta r^2} dl \quad (4)$$

10 Where:

11 $q_{p,i}$ stands for source strength of pollutant p , which is determined by its emission factor

12 $E_{f,p,i}$ and vehicle fraction s_i ,

13 r stands for the distance from the observed point,

14 β_1 is a coefficient.

15

16 The source strength is calculated by (23)

$$q_{p,i} = \frac{\sum s_i E_{f,p,i}}{3600} \quad (5)$$

$$E_{f,p,i} = \frac{\sum_k Q_{E,p,i,k}}{V_{KT,i}} \quad (6)$$

1 where:

2 $Q_{E,p,i,k}$ stands for the quantity of pollutant p emitted by vehicle k on link segment i ;

3 $\sum_k Q_{E,p,i,k}$ stands for the total emission of pollutant p on link segment i considering all
4 different types of vehicles

5 $V_{KT,i}$ stands for the activities (vehicle-kilometer traveled) associated with pollutions on
6 link segment i .

7 • **Unit Intake Damage Cost**

8 As the key cost component of travel, emission is more likely to be categorized into exter-
9 nal cost, which results in the lack of assessment of unit intake damage cost. Considering
10 the correlations between the emitted and breathed in emissions, the unit intake damage
11 cost could be estimated by,

$$u_{I,p} = \frac{u_{E,p}}{F_I} \quad (7)$$

12 Where:

13 $u_{E,p}$ stands for the unit emission cost of pollutant p ;

14 F_I stands for the intake fraction, which is the fraction of emissions that are taken in by
15 people (25).

16 National Highway Traffic Safety Administration (26) estimated the unit emission cost
17 ($u_{E,p}$), which shows the values of reductions in health damage costs per ton of emission
18 of each pollutant that is avoided. Assuming that the intake fraction is 10 per million (25),
19 the unit cost of both intake-emission and emission of each pollutant is showing in Table
20 3.

TABLE 3 : Unit Cost of Emission Intake and Emissions

	Unit Intake-Emission Cost (\$/g)	Unit Emission Cost (\$/ton)
Nitrogen oxides	670	6,700
Particulate matter	30,650	306,500
Sulfur dioxide	3,960	39,600
Carbon dioxide	0	22

¹ CO_2 has external effects on climate change, but no internal effect on health damage,

² Unit emission cost of CO_2 varies with different years, \$22 is the value of 2010.

1 External Emission Cost Estimation

2 The external emission cost from the perspective of travelers refers to the health damage cost from
3 emitted pollutants imposing on others and climate change cost due to CO_2 .

4 Two types of external emission cost would be considered, average external emission cost
5 ($E_{C,E,Avg}$) and marginal external emission cost ($E_{C,E,Mar}$), in which the marginal cost represent the
6 increased external emission cost due to an additional vehicle on the road.

$$E_{C,E,Total} = \sum_p Q_{E,p} * u_{E,p} \quad (8)$$

$$E_{C,E,Avg} = \frac{E_{C,E,TotalAuto}}{P_{Auto}} \quad (9)$$

$$E_{C,E,Mar} = \frac{\partial E_{C,E,Total}}{\partial P_{Auto}} \quad (10)$$

9 Where:

10 $E_{C,E,Total}$ stands for the total external emission cost, $E_{C,E,TotalAuto}$ is specific to the external emission
11 cost generated by auto;

12 P stands for the population, P_{Auto} refers to the population of auto;

13 $Q_{E,p}$ stands for the quantity of emission for pollutant p ;

14 $u_{E,p}$ stands for the unit emission cost of pollutant p ;

15 In practice, to reduce simulation noise, we measured the incremental emission cost instead
16 of the marginal cost by simulating 100 more vehicles than observed, calculating the increase of
17 emission cost, and dividing by 100. The expression of incremental external emission cost could be
18 written as:

$$E_{C,E,Inc} = \frac{\Delta E_{C,E,Total}}{\Delta P_{Auto}} \quad (11)$$

19 The Greenest Path

20 The greenest path, the route with the lowest emission cost, was proposed as a new rule of traffic
21 route assignment to minimize the emission cost. From both the internal and external cost versions,
22 greenest (internal) Path ($P_{g,i}$) pursues an Environmentally User Optimal (EUO) to minimize the
23 on-road intake-emission. A complement to this, greenest (external) Path ($P_{g,e}$) pursues an Environ-
24 mentally System Optimal (ESO) considering the emissions.

25 For a given OD pair, the greenest path could be expressed as,

$$E_{C,OD,k} = \sum_{i \in P_{OD,k}} E_{C,i} \quad (12)$$

$$E_{C,OD} = \min(E_{C,OD,k}) \quad (13)$$

26 Where:

27 $E_{C,i}$ stands for the emission cost of link segment i . In this case, it could be the internal or external
28 (average/incremental) emission cost,

29 $P_{OD,k}$ stands for the k^{th} path between origin O and destination D ;

- 1 $E_{C,OD,k}$ stands for the emission cost of the k^{th} path between O and D ;
 2 $E_{C,OD}$ stands for the minimum emission cost between O and D .
 3 The route with an emission cost of $E_{C,OD}$ is the greenest path as we defined.

4 **Betweenness**

5 Betweenness, as a network structure measure, was defined by Freeman (27) that quantifies the
 6 times of a link on a roadway network that is passed by particular paths between OD pairs. In our
 7 study, these paths refer to either the greenest path or the shortest travel time path. Its expression is
 8 shown as,

$$B_i = \frac{q_i}{Q} \quad (14)$$

9 Where:

- 10 B_i stands for the betweenness of link segment i ,
 11 q_i stands for the number of trips passing through i by using the shortest path,
 12 Q stands for the total number of trips.

13 Depending on the OD table of LEHD data, the total number of trips (Q) in Twin Cities (both
 14 origin and destination are located in the Seven County Metro area) is fixed for both the greenest
 15 path and the shortest travel time path. The changes of q_i for using the greenest path ($q_{G,i}$) and the
 16 shortest travel time path ($q_{T,i}$) are the parts that we are more concerned about, which we called the
 17 work trip flows.

18 **Cumulative Opportunity Measure**

19 Accessibility measures the ease of reaching activities (28). It combines the correlations between
 20 the cost and benefit of travel into a single metric, which represents the strengths and weaknesses
 21 of the interactions of transport network and land use, that makes accessibility a reliable tool for
 22 comparing the effectiveness of proposed land-use and transport network scenarios in planning
 23 projects (29, 30).

24 Various methods have been proposed to measure accessibility. A very basic one is the Cu-
 25 mulative Opportunity Measure that counts the number of opportunities that can be reached within a
 26 given threshold (31, 32). Comparing with other measures, Gravity-based Measure (28, 33), Utility-
 27 Based Measure (34), e.g., Cumulative Opportunity Measure is more applicable on the metropolitan
 28 area scale to a more straightforward cost function. This measure does generate artificial distinctions
 29 among destinations based on its binary cost function (35, 36), but a time-weighted accessibility,
 30 which combines the cost impedance with the measure, mitigates the distinctions (30, 37).

31 Most commonly, accessibility metrics have been analyzed from the perspective of the mean
 32 or expected travel time by considering the time cost of travel (38). In this study, accessibility
 33 measurement from the aspects of emission cost (internal and external) would be focused on.

34 The cumulative opportunity measure for job accessibility is typically expressed as:

$$A_O = \sum_D O_D f(W_{OD}) \quad (15)$$

35

$$f(TC_{OD}) = \begin{cases} 1 & \text{if } C_{OD} \leq W \\ 0 & \text{if } C_{OD} > W \end{cases} \quad (16)$$

1 Where:

2 A_O stands for the job accessibility of origin O ,

3 O_D stands for the number of jobs in destination D ,

4 C_{OD} stands for the travel costs between origin O and destination D . From the perspective of
5 emission cost, C_{OD} refers to the $E_{C,OD}$ (minimum emission cost). While from the perspective of
6 time cost, it refers to the $T_{C,OD}$ (minimum time cost),

7 W represents travelers' willingness-to-pay for corresponding travel costs (thresholds).

8 Moreover, accessibility loss was proposed to express the penalties to pursue the optimal
9 path from one aspect rather than another in terms of the reduction of reachable destinations with
10 the same willingness-to-pay. For instance, the accessibility loss of using the greenest path from the
11 perspective of time cost (A_{C_E,O,W_T}) is calculated by,

$$A_{C_E,O,W_T} = A_{O,T,W_T} - A_{O,E,W_T} \quad (17)$$

12 Where:

13 A_{O,T,W_T} stands for the accessibility of origin O based on the shortest travel time path within a time
14 threshold of W_T ,

15 A_{O,E,W_T} stands for that of using the greenest path within the same time threshold.

16 RESULTS

17 Internal vs. External Emission Cost

18 The mean value of internal emission cost of link segments in the Twin Cities is approximately
19 \$0.0025. Most (90%) link segments have an internal emission cost less than \$0.010. External
20 emission costs are much higher than internal costs, the mean value of the external emission aver-
21 age cost is \$0.0064 and incremental cost is around \$0.0066 (It is reasonable that the incremental
22 external emission cost is slightly larger than the average one since congestion affects the incremen-
23 tal emission cost more significantly). About 90% of link segments have an external emission cost
24 below \$0.015. Figure 2 shows the patterns of on-road emission cost per kilometer traveled.

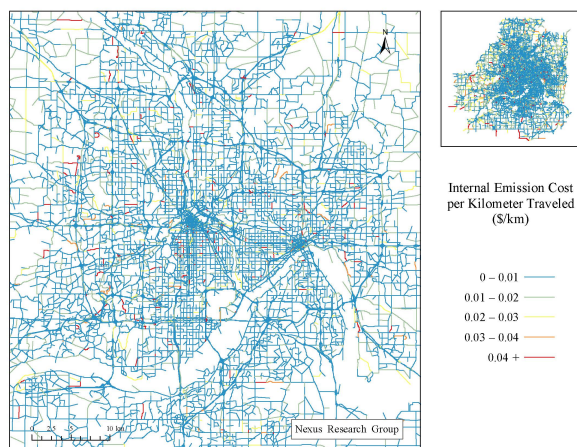
25 Figure 2 indicates a similar pattern, the external emission cost is higher than the internal
26 one. It implies that the emission costs travelers impose on others are greater than those borne by
27 themselves, which is to be expected, as the external unit costs include damage to non-travelers,
28 while the internal costs here exclude pollution costs from non-transport sources.

29 Work Trip Flow Estimation

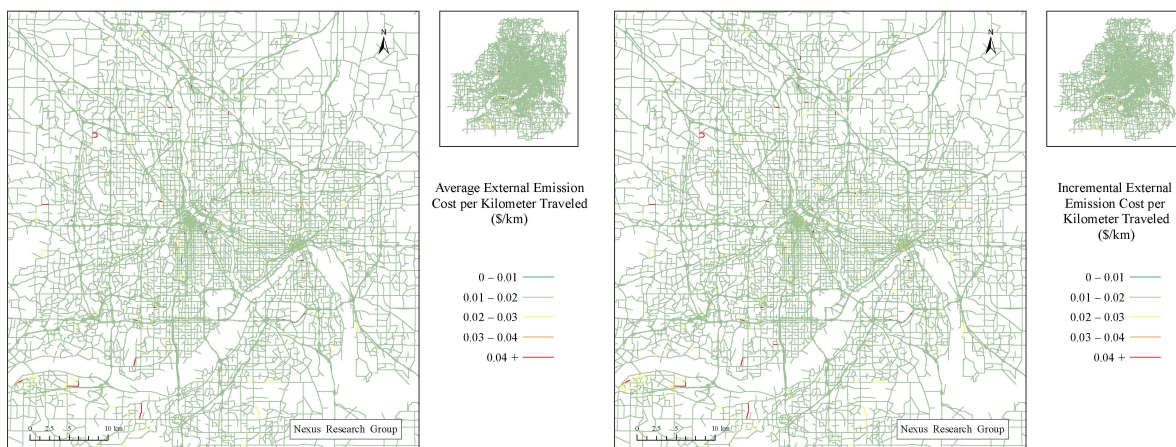
30 Figure 3 shows the work trip flow if all workers in the LODS data followed the greenest path be-
31 tween home and work, and the resulting pollution were unchanged (i.e. this is a partial equilibrium
32 analysis, we are not recomputing pollution costs based on the resulting traffic flows).

33 Generally, using the greenest internal path (Figure 3a), highways carry more traffic than
34 local roads, which is consistent with shortest path analysis (and observed data), since they are
35 faster, travelers have less exposure to pollution.

36 Considering the optimization of external emission cost, Figure 3b and 3c shows the work
37 trip flows based on the greenest (external) path considering average and incremental cost respec-
38 tively. Generally for greenest external path, the routes which are selected are more direct and thus
39 have shorter distances over which to emit pollutants. These differ from Figure 3a in that the trips
40 are more evenly distributed on the highway network except some critical links or interchanges,
41 such as the interchange of I-94 and I-35W, considering the external emission cost.



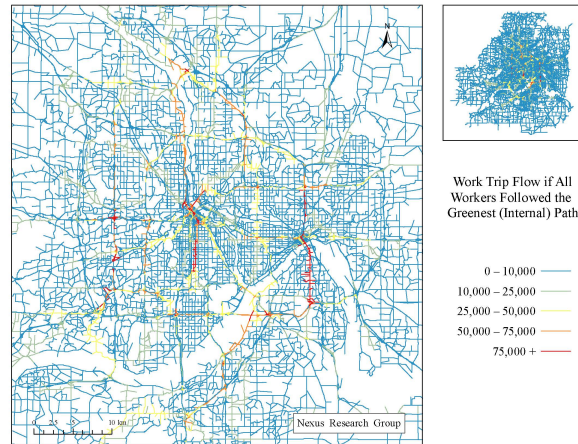
(a) Internal Emission Cost



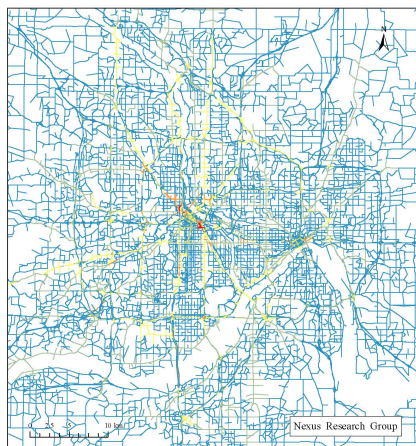
(b) External (Average) Emission Cost

(c) External (Incremental) Emission Cost

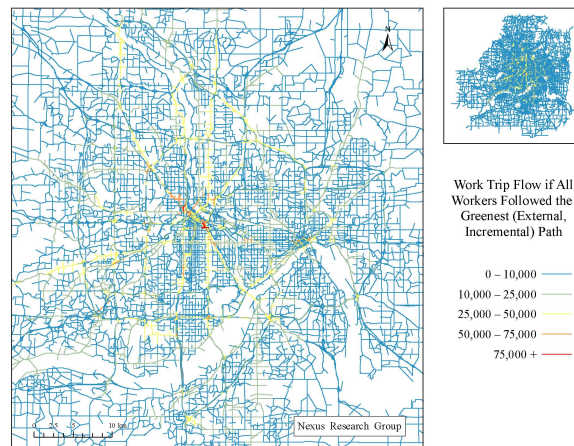
FIGURE 2 : Emission Cost Per Kilometer Traveled



(a) Greenest (Internal) Path



(b) Greenest (External, Average) Path



(c) Greenest (External, Incremental) Path

FIGURE 3 : Work Trip Flows

1 The work trip flows based on the average and incremental greenest (external) path do not
 2 have critical differences, consistent with the results shown in Figure 2

3 **Accessibility Measurement**

4 *Greenest (Internal) Path*

5 Job accessibility based on the greenest (internal) path are shown in Figure 4. The intake emission
 6 cost threshold (W) changes from \$0.05 to \$0.30.

7 The basic distribution pattern of job accessibility in the realm of internal emission cost is
 8 expected that the zones with higher job accessibility are centered on downtown Minneapolis (red
 9 color zones). In each figure, the accessibility declines gradually with the increase of distance to
 10 the downtown area, and the exurban area has the lowest job accessibility.

11 The major reason for such a pattern is that job opportunities are centered on downtown
 12 Minneapolis in the Twin Cities region. Residents living in suburban and exurban area require
 13 more time to reach the same number of job opportunities, which represents more exposure time
 14 for intake emissions.

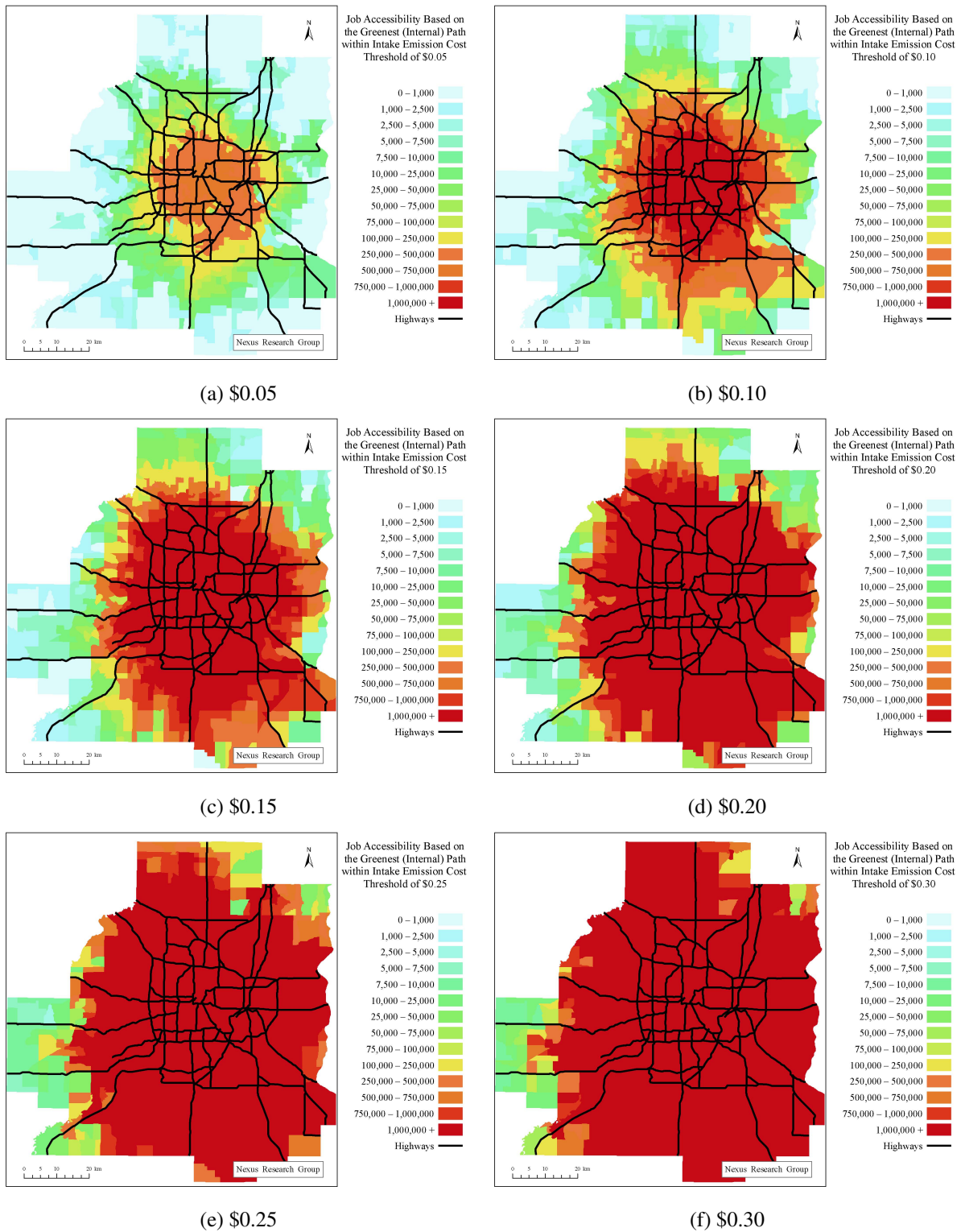


FIGURE 4 : Job Accessibility Based on the Greenest (Internal) Path with Different Intake-emission Cost Thresholds from \$0.05 to \$0.30

1 From Figure 4a to 4f, job accessibility increases significantly due to a higher level of cost
2 threshold. Within \$0.30 of internal emission cost, most of the region's residents can reach most
3 job opportunities. It reflects the effects of travelers' willingness-to-pay for intake-emission on
4 accessibility.

5 *Greenest (External) Path*

6 Using the greenest (external) path, the job accessibility within different cost thresholds (again from
7 \$0.05 to \$0.30) are displayed in Figure 5 and 6.

8 As expected, for both average and incremental external emission cost, accessibilities are
9 distributed as the similar pattern as Figure 4. However, within the same cost thresholds, the ac-
10 cessibilities are much lower when the greenest (external) path is pursued rather than the greenest
11 (internal) path. Such a result corresponds with the emission cost assessment that the external emis-
12 sion cost is higher than the intake cost.

13 Comparing Figure 5 and 6, accessibility based on the average external emission cost is
14 slightly larger than incremental cost. Figure 7 shows the differences directly. A clear ring of this
15 difference is shown on each map in Figure 7. The radius of the ring increases along with a higher
16 cost threshold.

17 **Accessibility Difference Analysis**

18 Figure 8 shows the accessibility loss of using the greenest (internal) path comparing to the tradi-
19 tional shortest travel time path.

20 Accessibility loss is higher in downtown Minneapolis with a lower willingness-to-pay for
21 time cost. It indicates that using the greenest (internal) path affects job accessibility for the resi-
22 dents living close to the downtown area the most if they have a lower willingness-to-pay for time
23 cost. The major reason is that residents living in or around downtown Minneapolis can reach more
24 opportunities in less time. While using the greenest (internal) path extends their travel time, say
25 more than 10 min, which results in a greater accessibility loss with a 10 min time threshold.

26 From Figure 8a to 8f, accessibility loss in the downtown area declines first with an increased
27 time threshold as the greenest path is by definition not as efficient from a time-perspective as the
28 shortest path, so jobs that were within 10 minutes on the shortest path are not within 10 minutes on
29 the greenest (internal) path. As time thresholds increase, the exurban areas are more affected, as
30 jobs that were within, say, 60 minutes of shortest path, are not within 60 minutes of greenest path.

31 **CONCLUSION**

32 Generally, on-road emissions are categorized as external cost expressing the health damage cost
33 from emitted pollutants imposed on others. However, as active agents in transportation system,
34 travelers bear health damage costs due to pollution intake, which logically should be considered as
35 an internal cost of travel.

36 Referring to the traditional shortest travel time path, the greenest (internal/external) paths
37 were proposed to estimate the minimum pollution exposure and emissions costs during traveling.
38 The greenest path presents the optimal solution of route choices for travelers (internal) or society
39 (external) from the perspective of pollution costs.

40 This study analyzed the internal and external emission cost on the scale of road network in
41 the metropolitan area, and evaluated the job accessibility for the Minneapolis-St.Paul region in the
42 realm of emission cost based on the greenest (internal or external) path.

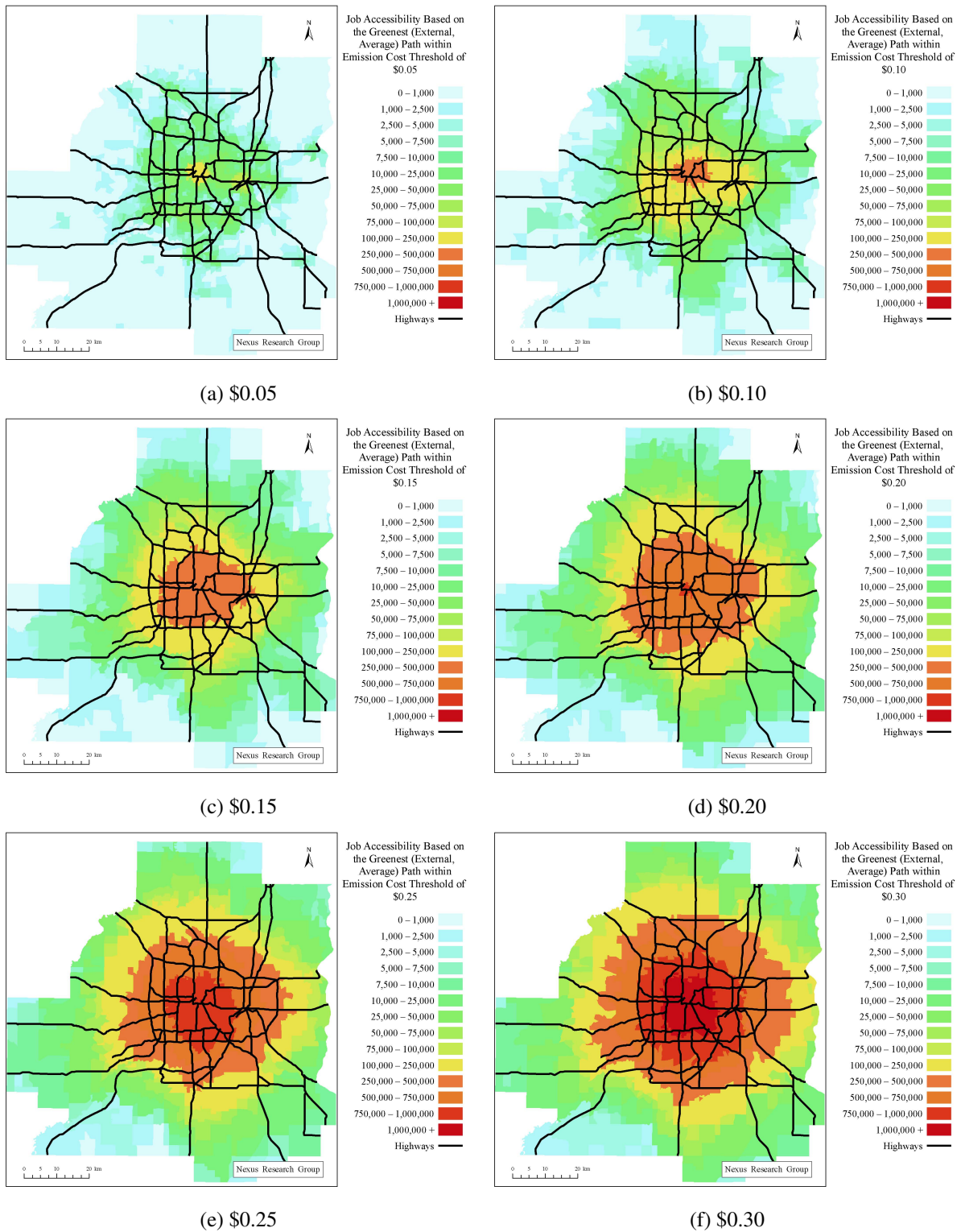


FIGURE 5 : Job Accessibility Based on the Greenest (External) Path (Average) with Cost Threshold from \$0.05 to \$0.30

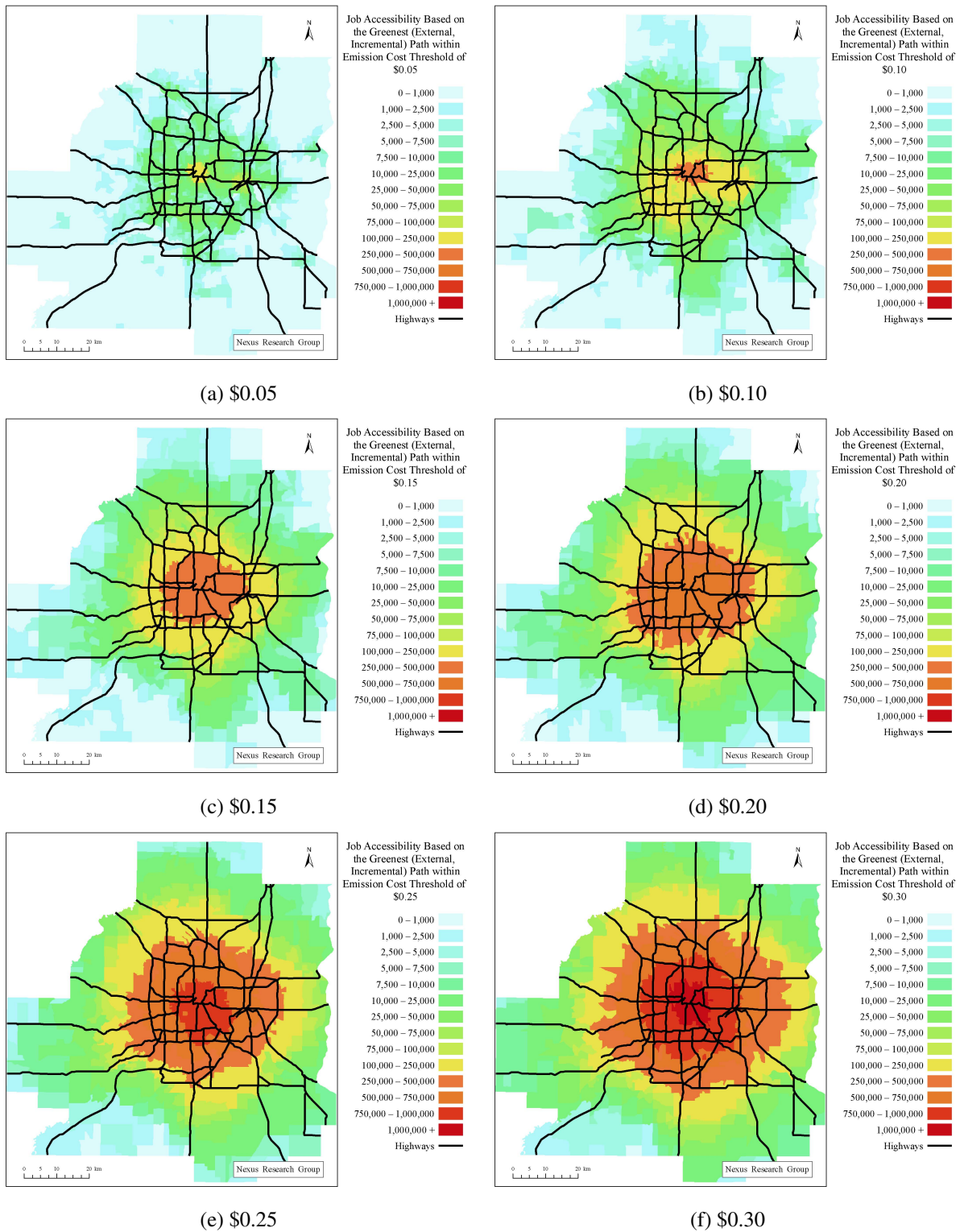


FIGURE 6 : Job Accessibility Based on the Greenest (External) Path (Incremental) with Cost Threshold from \$0.05 to \$0.30

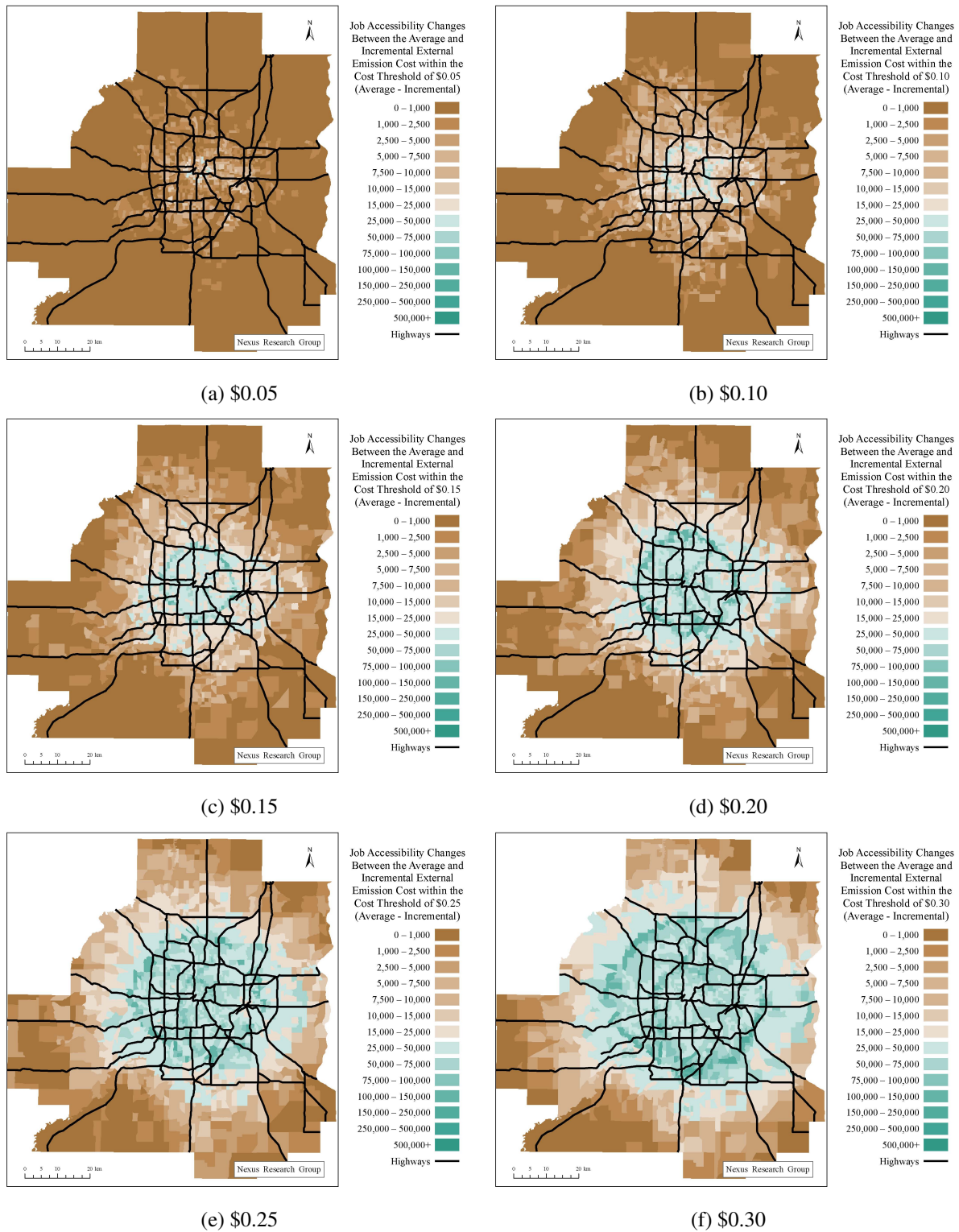


FIGURE 7 : Differences of Accessibility between the Average External Emission Cost and the Incremental External Emission Cost with Cost Threshold from \$0.05 to \$0.30

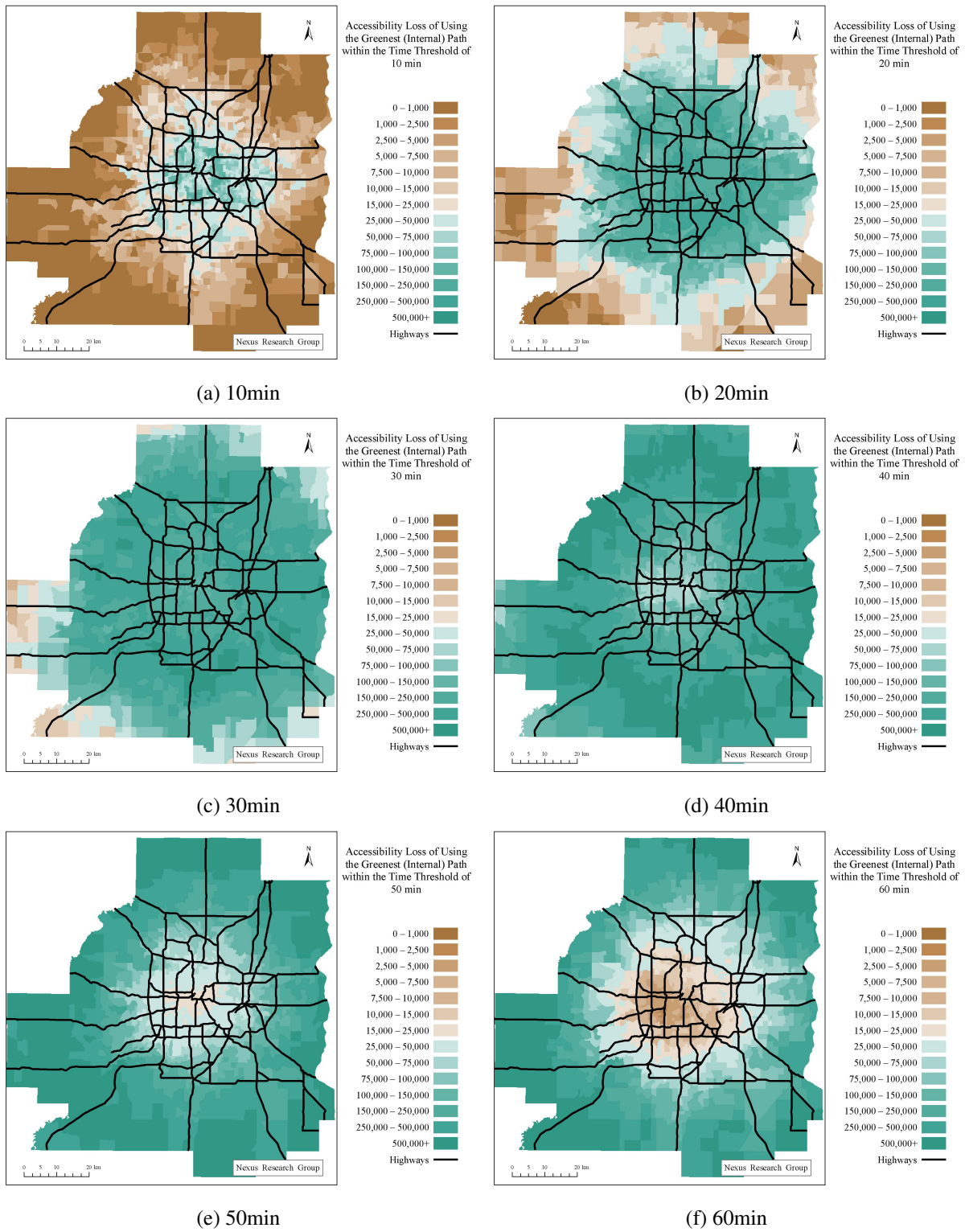


FIGURE 8 : Accessibility Loss of Using the Greenest Path Comparing with the Shortest Path with Different Time Threshold from 10min to 60min

1 Depending on the cost analysis, basically, the emission cost travelers impose on others
2 (external) is greater than that borne by themselves (internal). The average external emission cost
3 is slightly lower than the incremental one.

4 The environmental cost accessibility metrics show similar patterns as traditional travel time
5 based measures – downtown Minneapolis has a higher job accessibility. Accessibility decreases
6 along with an increase of the distance to downtown. Exurban areas have the lowest job accessi-
7 bility under all conditions. Emission cost threshold, for both internal and external costs, affects
8 accessibility as a higher emission cost threshold results in a higher job accessibility overall.

9 Job accessibility based on the greenest (internal) path is much greater than that based on the
10 greenest (external) path, which is consistent with emission cost analysis. Comparing externality
11 measures, the accessibility based on the average emissions cost is slightly larger than incremental
12 emissions cost, and their differences generate a clear ring, which represents a greater difference,
13 on the maps. Using the greenest path illustrates a higher accessibility loss in downtown Minneapo-
14 lis when there is a lower willingness-to-pay for time cost. It declines with the increase of time
15 threshold and the accessibility compensation in the exurban area rises at the same time.

16 From a policy perspective, road pricing presents a family of potential mechanisms to en-
17 courage use of socially optimal routes. Present implementations of road pricing are quite crude
18 compared to what is technically feasible. Currently, prices are fixed by area (there is a fixed charge
19 to drive into central London, Singapore, or Stockholm for the day), or by link (e.g. most highway
20 or bridge tolls) or for a given on ramp - off ramp pair (e.g. the New Jersey Turnpike). There are
21 off-peak discounts on many priced roads. Further, a few facilities vary by time of day (e.g. SR
22 91 in southern California) or dynamically (e.g. the High Occupancy/Toll lanes on I-394 in Min-
23 neapolis). However the technology exists to geolocate individual vehicles and charge tolls varying
24 by time of day, and by the specific route chosen to connect the origin and destination, and thus by
25 the level of pollution produced or inadvertently consumed.

26 Future research should consider a full cost analysis of other key cost components of travel
27 and evaluate the accessibility from the perspective of alternate cost components, time, safety,
28 money, e.g.. Moreover, a full cost model should combine those elements together, including both
29 internal and external cost, and measure the accessibility based on the lowest internal cost and ex-
30 ternal cost path. The accessibility by alternate traffic modes based on travel cost analysis will also
31 be measured.

32 REFERENCES

- 33 [1] Dockery, D., Epidemiologic evidence of cardiovascular effects of particulate air pollution.
34 *Environmental health perspectives*, Vol. 109, No. Suppl 4, 2001, p. 483.
- 35 [2] Pope III, C., R. Burnett, M. Thun, E. Calle, D. Krewski, K. Ito, and G. Thurston, Lung
36 cancer, cardiopulmonary mortality, and long-term exposure to fine particulate air pollution.
37 *Jama*, Vol. 287, No. 9, 2002, p. 1132.
- 38 [3] Gillies, J., A. Gertler, J. Sagebiel, and W. Dippel, On-road particulate matter (PM_{2.5} and
39 PM₁₀) emissions in the Sepulveda Tunnel, Los Angeles, California. *Environ. Sci. Technol.*,
40 Vol. 35, No. 6, 2001, pp. 1054–1063.
- 41 [4] Charron, A. and R. Harrison, Fine (PM_{2.5}) and coarse (PM_{2.5-10}) particulate matter on a

- 1 heavily trafficked London highway: Sources and processes. *Environ. Sci. Technol.*, Vol. 39,
2 No. 20, 2005, pp. 7768–7776.
- 3 [5] Lee, P., J. Brook, E. Dabek-Zlotorzynska, and S. Mabury, Identification of the major sources
4 contributing to PM_{2.5} observed in Toronto. *Environ. Sci. Technol.*, Vol. 37, No. 21, 2003, pp.
5 4831–4840.
- 6 [6] Donaldson, K., X. Li, and W. MacNee, Ultrafine (nanometre) particle mediated lung injury.
7 *Journal of Aerosol Science*, Vol. 29, No. 5-6, 1998, pp. 553–560.
- 8 [7] Warheit, D., T. Webb, C. Sayes, V. Colvin, and K. Reed, Pulmonary instillation studies with
9 nanoscale TiO₂ rods and dots in rats: toxicity is not dependent upon particle size and surface
10 area. *Toxicological sciences*, Vol. 91, No. 1, 2006, p. 227.
- 11 [8] Jaques, P. and C. Kim, Measurement of total lung deposition of inhaled ultrafine particles in
12 healthy men and women. *Inhalation Toxicology*, Vol. 12, No. 8, 2000, pp. 715–731.
- 13 [9] Oberdörster, G., E. Oberdörster, and J. Oberdörster, Nanotoxicology: an emerging disci-
14 pline evolving from studies of ultrafine particles. *Environmental health perspectives*, Vol.
15 113, No. 7, 2005, p. 823.
- 16 [10] Oberdörster, G., Z. Sharp, V. Atudorei, A. Elder, R. Gelein, W. Kreyling, and C. Cox, Translo-
17 cation of inhaled ultrafine particles to the brain. *Inhalation Toxicology*, Vol. 16, No. 6-7, 2004,
18 pp. 437–445.
- 19 [11] Mayeres, I., S. Ochelen, and S. Proost, The marginal external costs of urban transport. *Trans-
20 portation Research Part D: Transport and Environment*, Vol. 1, No. 2, 1996, pp. 111–130.
- 21 [12] Small, K. A. and C. Kazimi, On the costs of air pollution from motor vehicles. *Journal of
22 Transport Economics and policy*, 1995, pp. 7–32.
- 23 [13] Ahn, K. and H. Rakha, Field Evaluation of Energy and Environmental Impacts of Driver
24 Route Choice Decisions. In *Intelligent Transportation Systems Conference, 2007. ITSC 2007*,
25 IEEE, 2007, pp. 730–735.
- 26 [14] Lena, T., V. Ochieng, M. Carter, J. Holguín-Veras, and P. Kinney, Elemental Carbon and
27 PM 2.5 Levels in an Urban Community Heavily Impacted by Truck Traffic. *Environmental
28 Health Perspectives*, Vol. 110, No. 10, 2002, pp. 1009–1016.
- 29 [15] TomTom International BV, *Speed Profiles*, 2013.
- 30 [16] Metropolitan Council, *Transportation analysis zones 2010*, 2012.
- 31 [17] US Census Bureau, *LEHD origin-destination employment statistics dataset structure format
32 version 7.0.*, 2013.
- 33 [18] United States Environmental Protection Agency, *Motor Vehicle Emission Simulator*, 2016.
- 34 [19] Minnesota Department of Transportation, *IRIS Software Distribution Home*, 2014.

- 1 [20] Wilde, W. J. and T. J. Stahl, *Update of Vehicle Classification for County Road Pavement*
2 *Design*, 2010.
- 3 [21] Traffic Forecasts and Analysis Section, MnDOT office of Transportation Data and Analysis,
4 *MnDOT Procedure Manual for Forecasting Traffic on Minnesota's Highway Systems*, 2012.
- 5 [22] Hassanien, M. A., N. M. Abdellatif, E. A. Saleh, and A. M. Mohamed, Inhalation Intake
6 Assessment of Air Pollutants Exposure Over Cairo, Egypt. In *Exposure and Risk Assessment*
7 *of Chemical Pollution—Contemporary Methodology*, Springer, 2009, pp. 303–315.
- 8 [23] Jin, T. and L. Fu, Application of GIS to modified models of vehicle emission dispersion.
9 *Atmospheric Environment*, Vol. 39, No. 34, 2005, pp. 6326–6333.
- 10 [24] Jiang, C. W. J. R., W.M., *Air Pollution Meteorology*. Meteorology Publishing Company, 1993.
- 11 [25] Marshall, J. D., S.-K. Teoh, and W. W. Nazaroff, Intake fraction of nonreactive vehicle emis-
12 sions in US urban areas. *Atmospheric Environment*, Vol. 39, No. 7, 2005, pp. 1363–1371.
- 13 [26] National Highway Traffic Safety Administration, Final Regulatory Impact Analysis: Corpo-
14 rate Average Fuel Economy for MY 2017-MY 2025 Passenger Cars and Light Trucks, 2010.
- 15 [27] Freeman, L. C., A set of measures of centrality based on betweenness. *Sociometry*, 1977, pp.
16 35–41.
- 17 [28] Hansen, W. G., How accessibility shapes land use. *Journal of the American Institute of plan-*
18 *ners*, Vol. 25, No. 2, 1959, pp. 73–76.
- 19 [29] Owen, A. and M. Kadziolka, *Green Line LRT: Job Accessibility Impacts in Minneapolis and*
20 *Saint Paul*, 2015.
- 21 [30] Anderson, P., D. Levinson, and P. Parthasarathi, Accessibility futures. *Transactions in GIS*,
22 Vol. 17, No. 5, 2013, pp. 683–705.
- 23 [31] Vickerman, R. W., Accessibility, attraction, and potential: a review of some concepts and
24 their use in determining mobility. *Environment and Planning*, 1974.
- 25 [32] Wachs, M. and T. G. Kumagai, Physical accessibility as a social indicator. *Socio-Economic*
26 *Planning Sciences*, 1973.
- 27 [33] Handy, S. L. and D. A. Niemeier, Measuring accessibility: an exploration of issues and alter-
28 natives. *Environment and planning A*, Vol. 29, No. 7, 1997, pp. 1175–1194.
- 29 [34] Ben-Akiva, M. E. and S. R. Lerman, *Discrete choice analysis: theory and application to*
30 *travel demand*, Vol. 9. MIT press, 1985.
- 31 [35] Owen, A. and D. Levinson, *Access to destinations: Annual accessibility measure for the twin*
32 *cities metropolitan area*, 2012.
- 33 [36] El-Geneidy, A. M. and D. M. Levinson, *Access to destinations: Development of accessibility*
34 *measures*, 2006.

- 1 [37] Levinson, D. M. and A. Kumar, A multi-modal trip distribution model. *Transportation Re-*
- 2 *search Record*, Vol. 1446, 1994, pp. 124–131.
- 3 [38] Cui, M. and D. Levinson, *Accessibility and the Ring of Unreliability*, 2016.